

Environmental Life Cycle Assessment (LCA) Of Wheat Cultivation In Agro industry company Of Iranian Novin(Aq-Qala)

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ABSTRACT

The unique and systematic solution to the problems existing in the energy and environmental sectors of Iran is to improve energy productivity (EP) in various fields. Despite the extensive researches on the energy consumption and optimization of crops in Iran and other countries, the direct study of pollution and greenhouse gas emissions has received less attention from researchers. The study of crop production from the viewpoint of environmental issues and emissions will be a solution to the environmental protection and achievement of sustainable agriculture. In this study, environmental pollution and energy input/output were investigated in Agroindustry Company of Iranian Novin Farms, Golestan province. According to the results, the average of energy output to input ratio was 1.04, with average EP of 0.07 kg/MJ, in wheat production. The average energy intensity and net energy gain were 14.1 MJ/kg and 1298.3 MJ/ha, respectively. Nursing and tillage operations accounted for 57 and 24% of total energy input, respectively. The results of wheat LCA using SIMAPRO software revealed that the open water poisoning index was 55845.08 kg 1,4-DCB eq. for the production of one ton of wheat. The global warming potential was 841 kg CO₂ due to diesel fuel consumption. Despite the negligible share of chemical pesticides from total energy consumption, a value of 257.3 kg was obtained for the human toxicity index per ton of the produced crop.

Keywords: Diesel Fuel, Energy, Life Cycle Assessment, Pollution, Wheat

INTRODUCTION

Environmental issues have received increasing attention in recent years. With increasing environmental awareness, various economic sectors have focused on evaluating the environmental impact of their activities. High use of energy by human causes many problems for the environment and human health. The increase in greenhouse gases raises global temperature and climate change problems. An accurate and comprehensive assessment that can simultaneously encompass all

environmental impacts requires the use of integrated criteria. Life cycle assessment or analysis (LCA) is a method in which all environmental impacts related to a product (including goods and services) are evaluated in the whole life cycle of that product from the extraction and/or collection of raw materials to the consumption stage and then recycling or removing the resultant wastes. Environmental assessment is a systematic and multistage process consisting of four stages, viz. goal and scope definition, inventory analysis, impact assessment, and interpretation (ISO 14040, 2016). Given the effectiveness of LCA, it is introduced as the most appropriate method for the sustainability assessment of various activities, including agriculture.

Agriculture is one of the important economic sectors that has important effects on the environment. Agriculture is responsive to food security, therefore, the increasing world population requires more agricultural activity. In the past decades, this sector is associated with the dramatically increased use of agricultural mechanization, widespread application of fertilizers and pesticides, and advances in animal husbandry, resulting in elevated energy input to farms (Roder et al., 2014). Agriculture is the main source of several important environmental issues such as eutrophication open waters poisoning and etc. (Zhu et al, 2018, Dijkman et al., 2018).

Optimal use of energy in the agricultural systems can reduce input energy in food production and pave the way towards sustainable agricultural production (Fernanda et al., 2016). Any agricultural activity will be ecologically sustainable when the amount of depleted resources and emitted pollutions from this activity could be compensated by nature. The environmental impacts of agricultural activities should be initially evaluated by considering their sustainability in the long term (Kaab et al, 2019, Borghino et al., 2021).

Borghino et al., (2021) investigated environmental problems caused by the use of different levels of nitrogen fertilizers in wheat production in Germany. They found that eutrophication was increased with rising nitrogen fertilizer use, which is one of the most important environmental impacts of wheat production. Their results indicated that acidity and global warming indices are other major environmental impacts in wheat production. Wowra et al. (2020) reported that changing the tillage method to which led to less carbon decomposition in agricultural soils, would reduce greenhouse gas emissions. In many studies, fossil fuel consumption has been reported as the most important factor for greenhouse gas emissions (Wang et al., 2020).

The use of pesticides and chemical fertilizers in agriculture is also the cause of emissions in the environment. Modification of pest control methods, calibration of sprayers, biological control against pests, and the use of integrated methods are effective ways in the reduction of pesticides consumption. Optimal use of chemical fertilizers and the use of biological fertilizers are effective in the reduction of emissions. According to data collected from farmers in the study of Singh et al.

(2021), only 25% of farmers used the integrated method (combined chemical and biological control) in their farms management.

Due to scarce studies on environmental effects of wheat production in Iran, this study intended to investigate the energy consumption in wheat cultivation stages and evaluate several environmental factors on wheat production in the farms of Iranian Novin Co.

MATERIALS AND METHODS

The geographical range of this research was wheat farms of Agroindustry Company of Iranian Novin (Aq-Qala in Golestan province, Iran). The statistics and information were obtained from Company's registered data for output-inputs and machinery operations in wheat crop cultivation. Overall data for climatic conditions, farms area, and land plots specifications were obtained by consulting with company's staffs.

Due to the large area under the cultivation by company, for better management the total farms area has been divided into four units. Each unit was with a somewhat different cropping method. Therefore, the data for this study were collected separately from each unit.

According to the ISO 14040 standard definition, LCA deals with environmental and potential environmental aspects throughout the life cycle of a crop production. Therefore, all input and output data in the form of consumed materials, direct energy consumption and machinery use were collected. Below, the four stages of LCA for this study is described.

Goal and scope definition: This study aimed to investigate the environmental impact groups affecting the depletion of water and fossil fuel resources, global warming, and acidification in wheat production. The functional unit was the production of one ton of wheat in Iranian Novin Co. Farms.

Inventory analysis:

The inputs were fuels, chemical pesticides, and fertilizers consumptions in wheat production farms. The calculation of energy input through main inputs was as below:

The actual capacity of the machinery operations is the area covered in one hour of machinery work. In some studies, an average coefficient is considered for all machines, which presents energy intensity in terms of machinery working hours (Moradi, 2018). This coefficient was used to obtain the amount of energy input from machines, knowing their working hours in each hectare. An average amount of fuel consumption was used to calculate the energy input from fuel consumption in different machinery operations. The production of chemical fertilizers is associated with the use of high energy levels. To calculate the energy input from chemical fertilizers, the used amount of different types of fertilizers consumed in different periods of wheat growth, was multiplied by their

energy equivalents. The same method was applied to calculate the energy input from animal manures and pesticides. The energy equivalents of these inputs in this study was according to the mentioned values by Fabiani(2020).

Emission levels emitted from wheat production farms are different depending on soil, climate, and farm management system, making it difficult to measure directly them in the farms. Hence, the estimations for emissions, such as the variety of nitrogen compounds emitted from the farms, were made according to the introduced factors in LCA studies (Borghino et al., (2021)).

Ammonia emission: Urea has the highest ammonia sublimation among mineral fertilizers. Ammonia sublimation level varies depending on climatic conditions and soil properties. Due to the lack of necessary studies in Iran in this study, the ammonia emission factor from urea was considered equal to the European and American averages. The sublimated $\text{NH}_3\text{-N}$ is about 17% of the total nitrogen consumed in the form of mineral fertilizers as well as urea (Borghino et al., (2021)).

N_2O emission: emitted N_2O into the atmosphere is affected by two microbial phenomena, viz. nitrification and denitrification. However, this issue is also influenced by different soil conditions, climate, and crop management. The factor presented by the International Panel on Climate Change (IPCC) was used in this study. According to the IPCC (2006), 1% of total consumed nitrogen is released as $\text{N}_2\text{O-N}$ (Ingram et al., 2018).

NO_x emissions: As reported by some investigations, emitted NO_x to the atmosphere was considered as 10% of N_2O levels.

Emissions from gasoline consumption: The most important types of greenhouse gases emissions, including CO_2 , N_2O , and CH_4 , from each liter gasoline combustion was calculated as follows (Naderi et al., 2020):

Impact assessment:

The stage of impact assessment aimed to further interpret the inputs and outputs in the wheat production system. Environmental profile was analyzed by LCA methodology tools using SimaPro 7.1 software (Pre Consultants, 2006), following the CML-2 baseline V3.01/EU25 method developed by the Institute of Environmental Sciences (CML) from Leiden University of the Netherlands. This model is commonly used in LCA studies of agricultural production. Finally, environmental indicators (global warming potential, acidification, eutrophication, toxicity, etc.) were identified. The calculated energy indices in crop production was as follows:

Energy output-input ratio (OIR) = Output Energy/Input Energy (1)

By definition, this indicator shows the status of energy outputs in relation to the input amounts.

$$\text{Net Energy Gain (NEG)} = \text{Output Energy} - \text{Input Energy} \quad (2)$$

By definition, this is the total amount of generated (output) energy minus the energy input (GJ/ha).

$$\text{Energy Productivity (EP)} = \text{Product yield} / \text{Energy input} \quad (3)$$

This is defined as the produced crop unit per energy input unit (kg/MJ). Energy productivity is considered as an important evaluation criterion in the comparison of energy efficiency of different production systems.

$$\text{Energy Intensity (EI)} = \text{the amount of consumed energy} / \text{the crop yield} \quad (11)$$

This index is defined as the energy amount consumed per unit of production (MJ/kg). It is the main indicator for comparison of energy efficiency of different cropping systems.

RESULTS AND DISCUSSION

LCA analyses

The calculated levels for environmental indicators in the studied wheat production are shown in Table 1. According to this table, for the production of one ton of wheat, 55845.1 kg of 1,4-DCB eq. was emitted, resulting the open water toxicity index (OWTI) as the main impact category. It is due to high CO₂ emissions from diesel fuel consumption. The second main impact was global warming potential (GWP) with 841 kg of CO₂ eq. emitted from N fertilizers in the form of N₂O. The human toxicity index (HTI) resulted from pesticides consumption was 257.3 kg of PO₄³⁻ eq. as the third main impact.

Table 1. Environmental impact indices in the wheat production

Effect section	Unit	Wheat
Depletion of organic resources	kg Sb eq.	3.4791
Acidification potential	kg SO ₂ eq.	5.1911
Eutrophication potential	kg PO ₄ ³⁻ eq.	0.9557
Global warming potential (GWP)	kg CO ₂ eq.	841.807
Ozone layer depletion	kg CFC-11 eq.	0.0003
Human toxicity potential	kg 1,4-DCB eq.	257.294
Surface water poisoning	kg 1,4-DCB eq.	21.0199
Open water poisoning	kg 1,4-DCB eq.	55845.08
Soil poisoning	kg 1,4-DCB eq.	0.3152
Photochemical oxidation potential	kg C ₂ H ₄ eq.	0.1349

The effects of chemical fertilizers, diesel fuel, and consumed pesticide are shown as a normalized diagram in Figure 1. Diesel fuel and consumed pesticides, with percentages of 60% and 39.8%, respectively, affected Surface water poisoning index. Among the substances emitted from diesel fuel, CO₂ with 80% had the greatest impact on the index.

All inputs had the same effect on the eutrophication index. Ammonium emitted from chemical fertilizers with a proportion of 72% had the highest effect on this index compared to N₂O and NO_x. Sulfur dioxide (SO₂) emitted from diesel fuel had the greatest impact on this index with 55% in comparison to other pollutants.

Diesel fuel and pesticides, with 61% and 38.5%, respectively, influenced soil toxicity index. Among the substances emitted from diesel fuel, CO₂ with 60% had the greatest effect on this index.

The most effective factors in the depletion of organic resources index were the use of diesel fuel and N fertilizers with 41% and 36%, respectively. Ammonium emitted from chemical fertilizers, with a proportion of 65%, had the greatest impact on this index compared to N₂O and NO_x. Among the substances emitted from diesel fuel, CO₂ with a proportion of 80% had the utmost impact on the index.

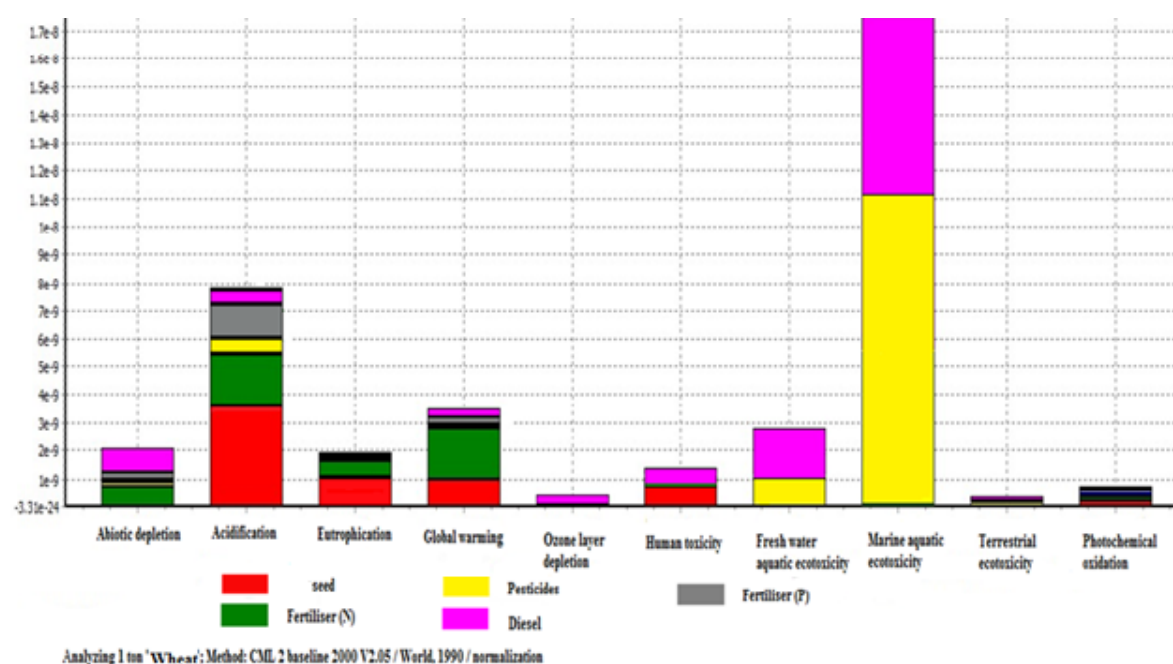


Figure 1. Levels of the environmental indices in the wheat production

The use of chemical fertilizers had no effect on the ozone depletion index while diesel fuel with 81% had a very strong effect on this index. CO₂ and CO emissions from the combustion of diesel fuel, each with a proportion of 40%, had the highest impact on this index.

Nitrate and phosphate fertilizers, with a share of 47% and 32%, respectively, had the greatest effect on the acidification potential (AP). N₂O emitted from chemical fertilizers with 85% had the uppermost

effect on this index. Among the materials emitted from diesel fuel, CO₂ with 75% had the greatest effect on AP.

The consumed inputs of pesticides and diesel fuel, with proportions of 51% and 46%, respectively, had the highest effect on the human toxicity index (HTI). Chromium with a 60% proportion had the greatest impact on this index compared to other diesel fuel emissions.

Chemical fertilizers, in particular nitrate fertilizer with a proportion of 76%, were the most effective factor on global warming potential (GWP) among the consumed inputs. N₂O emitted from chemical fertilizers, with 82%, had the highest impact on GWP compared to ammonium and NO_x. With a share of 90%, CO₂ emitted from diesel fuel had the uppermost effect on this index compared to CH₄, CO₂, N₂O, NO_x, and Non-methane volatile organic compounds (NMVOC).

Energy analyses

The total energy input to the system and the share of each input are determined given the energy equivalents to the construction and depreciation of used machinery and consumption of fuel, seeds, fertilizers, and chemical pesticides.

As shown in Table 3, a total energy of 30.8 GJ/ha was obtained for the wheat production in the company's farms. The highest energy consumption belongs to seed, fertilizers, and chemical pesticides, accounting for 60.01% of the total energy consumption. Chemical fertilizers are energy intensive, therefore, their higher consumption increases the energy input. According to this table, fuel energy input lies in the next category.

Table 3. The proportion of the inputs from total energy in wheat production (MJ/ha)

Machine	Energy	Percentage of total	Rank
Machinery	4637.7	15.1	3
Fuel	7677.5	24.9	2
Seed, fertilizers, and pesticides	18483.8	60.0	1
Total	30799.1	100.0	

According to Table 4, the output energy was 33.0 GJ/ha for wheat grain as the mean value for all farms. The farms in unit C had the highest yield compared to other units.

Khanali et al. (2018) reported the energy input and output of 19.3 and 19.4 GJ/ha for rain-fed wheat growing, where the output and input values were 33.0 and 31.7 GJ/ha in the present study. As reported by other study, in Golestan province energy input and output for the irrigated wheat production were 32.0 and 94.0 GJ/ha, representing higher crop yield in irrigated farms

(Kazemi and Zare, 2014). In the studied company the farms were rain-fed but the wheat yield was higher than other rain-fed systems in Iran (2887 kg/ha). However, it was lower than that for irrigated wheat cropping systems. In another energy study for rain-fed wheat in Mazandaran province, the energy input and output were 22.0 and 29.0 GJ/ha (Gholinejad and Hassanzadeh 2012).

Table 4. The energy output in the farm units for wheat production

Farm Unit	Yield (kg)	Energy content (MJ/kg)	Energy equivalent (MJ)	Rank
A	2150	14.7	31605	3
B	1904	14.7	27988.8	4
C	2327	14.7	34206.9	1
D	2137	14.7	31413.9	2
Weighted average	2244.4	14.7	32993.3	

Energy indices

According to the results of input and output energies (Table 3 and 4), energy indicators in wheat cultivation were calculated using the definitions provided for these indices.

According to Table 5, the energy output to input ratio (OIR) was 1.04 for all farms as the weighted average of the company units. Due to the higher yield of unit C, it can be observed that this unit has a better OIR (> 1), whereas this ratio is less than 1 for unit B due to lower yields, resulting in negative NEG for unit B. However, a positive net energy gain (NEG) index (1.3 GJ/ha) was calculated by weighted averaging in the whole company.

The OIR obtained for wheat in this study is lower than those of 3.13, 3.17, and 1.64 MJ/MJ reported by Afzalinia and Karimi (2020), Shahin et al. (2008) in Ardabil, Mashhoori Azar et al. (2007) in Maragheh, and Sadeghi (2008) in Mahyar plain of Shahreza, respectively. In the mentioned studies the farms were irrigated resulting in higher yields than this study. Compared to some rainfed cultivations, with ratios of 1.03, 0.68, 0.81, and 0.99 reported respectively for Shahrekord, West Azarbaijan, Kermanshah, and Saveh (Kazemi et al. 2016), the OIR in this study is somewhat higher.

In this study, the average energy productivity EP index (0.07 kg/MJ) is lower than that (0.096 kg/MJ) reported in irrigated wheat growing in Ardabil region (Shahin et al. 2008). A similar EP of 0.07 was obtained in rainfed wheat growing in Shahrekord region (Kazemi et al. 2016). The mean weighted energy intensity (EI) index was 14.12 MJ/kg for all farms, which is close to that of 15.67 for

rainfed wheat growing in Shahrekord. The calculated EI in this study is much higher than that of 8.21 for rainfed wheat in Kermanshah region (Fathi et al. 2018). In fact, it can be re-concluded many inputs were consumed to achieve a high rainfed yield in the Iranian Novin Co. farms. A review on a yield of over two tons per hectare indicates that it is much higher than the yields reported for rainfed farms, which are usually in the range of 1 t/ha in Iran.

Table 5. Energy indices in the wheat production units

	Unit	A	B	C	D	Total
OIR	MJ/MJ	1.01	0.87	1.01	1.09	1.04
EP	kg/MJ	0.07	0.07	0.07	0.07	0.07
EI	MJ/kg	14.60	16.92	14.61	13.51	14.12
NEG	MJ	225.00	-4229.20	207.90	2546.90	1298.27

Table 6 represents the energy consumed in the farms in individual stages of the crop cultivation. Accordingly, the highest amount of energy consumption (57%) belongs to the growing stage due to the high consumption of nitrogen fertilizers at this stage. If fertilizer use is saved at the growing stage, the amount of energy consumed will be reduced in this stage. The harvesting stage accounts for the lowest amount of energy consumption, which is less than 6% of the total input energy. Although the tillage stage usually has the highest energy consumption, it is in the second place in this company with 24% of energy consumption. The use of no-tillage or low-tillage methods will also decrease the energy consumption at this stage.

Table 6. The energy consumption in different stages of wheat cultivation

	Energy consumption (MJ/ha)	Energy consumption (%)
Tillage	7568.5	23.9
Cropping	4256.8	13.4
Growing	18041.5	56.9
Harvesting	1835.1	5.8

CONCLUSIONS

This study aimed to investigate the amount of in-farm emissions in wheat cultivation. Direct and indirect emissions from farms is due to the consumption of several inputs as seed, fertilizers, and fuels consumed in agricultural operations.

According to the results of LCA of wheat crop in the farms of Iranian Novin Agroindustry Co., marine aquatic eco-toxicity, acidification and global warming are three main environmental impact categories in the wheat production. Pesticides, fertilizers and diesel fuel are the main resources of the pollutants resulting in these impact categories. Given the energy input and output analysis of the farms, EI in wheat grain production was 14.12 MJ/kg (without considering straw as a byproduct). A comparison with other studies indicates that this value is much higher than that for irrigated wheat production but similar or less than that for other studies around rainfed wheat. The highest energy input in this study with 40% belongs to the fuel consumption followed by nitrogen fertilizers with 40%. To reduce the total energy input, alternative methods should be considered as well as using lower pesticides and fertilizers or application of reduced machinery operations without a significant reduction in wheat yield. Also, increasing the crop yield by irrigation program would reduce the energy intensity in wheat production and break down the emitted pollutions. The use of straw as a byproduct would also break down the energy input between product and byproduct.

REFERENCES

- Afzalinia, S., & Karimi, A. (2020). Barley Cultivars and Seed Rates Effects on Energy and Water Productivity of Green Fodder Production under Hydroponic Condition. *Indian Journal of Agricultural Research*, 54(6).
- Amienyo, D., & Azapagic, A. (2016). Life cycle environmental impacts and costs of beer production and consumption in the UK. *The international journal of life cycle assessment*, 21(4), 492-509.
- Borghino, N., Corson, M., Nitschelm, L., Wilfart, A., Fleuet, J., Moraine, M., ... & Godinot, O. (2021). Contribution of LCA to decision making: A scenario analysis in territorial agricultural production systems. *Journal of Environmental Management*, 287, 112288.
- Dijkman, T. J., Basset-Mens, C., Antón, A., & Núñez, M. (2018). LCA of food and agriculture. In *Life Cycle Assessment* (pp. 723-754). Springer, Cham.
- de Carvalho Macedo, I., Nassar, A. M., Cowiec, A. L., Seabra, J. E., & Marellid, L. (2015). Greenhouse gas emissions from bioenergy. *Bioenergy & Sustainability: Bridging the Gaps SCOPE*, 72, 582-617.
- Fabiani, S., Vanino, S., Napoli, R., & Nino, P. (2020). Water energy food nexus approach for sustainability assessment at farm level: An experience from an intensive agricultural area in central Italy. *Environmental Science & Policy*, 104, 1-12.

- Fernanda Lare-Orozco, M., Robles-Morua, A., Yepez, E., & Handler, R. (2016). Global warming potential of intensive wheat production in the Yaqui Valley, Mexico: a resource for the design of localized mitigation strategies. *Journal of Cleaner Production*, 1–11).
- Ingram, D. L., Hall, C. R., & Knight, J. (2018). Global warming potential, variable costs, and water use of a model greenhouse production system for 11.4-cm annual plants using life cycle assessment. *HortScience*, 53(4), 441-444.
- ISO14040. (2006). Environmental management — Life cycle assessment — Principles and framework. In. Geneva, Switzerland: International Organization for Standardization.
- ISO14044. (2006). Environmental management — Life cycle assessment — Requirements and guidelines. In. Geneva, Switzerland: International Organization for Standardization.
- Kaab, A., Sharifi, M., Mobli, H., Nabavi-Pelesaraei, A., & Chau, K. W. (2019). Combined life cycle assessment and artificial intelligence for prediction of output energy and environmental impacts of sugarcane production. *Science of the Total Environment*, 664, 1005-1019.
- Khanali, M., Mousavi, S. A., Sharifi, M., Nasab, F. K., & Chau, K. W. (2018). Life cycle assessment of canola edible oil production in Iran: a case study in Isfahan province. *Journal of Cleaner Production*, 196, 714-725.
- Moradi, M., Nematollahi, M. A., Khaneghah, A. M., Pishgar-Komleh, S. H., & Rajabi, M. R. (2018). Comparison of energy consumption of wheat production in conservation and conventional agriculture using DEA. *Environmental Science and Pollution Research*, 25(35), 35200-35209.
- Naderi, S., Raini, M. G. N., & Taki, M. (2020). Measuring the energy and environmental indices for apple (production and storage) by life cycle assessment (case study: Semirom county, Isfahan, Iran). *Environmental and Sustainability Indicators*, 6, 100034.
- Roder, M., Thornley, P., Campbell, G., & Bows-Larkin, A. (2014). Emissions associated with meeting the future global wheat demand: A case study of UK production under climate change constraints. *Environmental Science & Policy*, 39, 13–24.
- Shahin, S. Jafari, A. Mobli, H. Rafiee, S. Karimi, M. (2008). Energy use and economical analysis of wheat production in Iran: A case study from Ardabil province. *J. Agric. Techn*, 4, 77–88.
- Singh, P., Benbi, D. K., & Verma, G. (2021). Nutrient Management Impacts on Nutrient Use Efficiency and Energy, Carbon, and Net Ecosystem Economic Budget of a Rice–Wheat Cropping System in Northwestern India. *Journal of Soil Science and Plant Nutrition*, 21(1), 559-577.
- Wang, J., Zhang, L., He, X., Zhang, Y., Wan, Y., Duan, S., ... & Shi, X. (2020). Environmental mitigation potential by improved nutrient managements in pear (*Pyruspyrifolia* L.) orchards based on life cycle assessment: A case study in the North China Plain. *Journal of Cleaner Production*, 262, 121273.

- Wowra, K., Zeller, V., & Schebek, L. (2020). Nitrogen in Life Cycle Assessment (LCA) of agricultural crop production systems: Comparative analysis of regionalization approaches. *Science of the Total Environment*, 143009.
- Zhu, Z., Jia, Z., Peng, L., Chen, Q., He, L., Jiang, Y., & Ge, S. (2018). Life cycle assessment of conventional and organic apple production systems in China. *Journal of Cleaner Production*, 201, 156-168.